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Russell, Bayden D.; Harley, Christopher D. G.; Wernberg, T.; Mieszkowska, Nova; Widdicombe, Stephen; Hall-Spencer, Jason M.; Connell, Sean Duncan

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3 Predicting ecosystem shifts requires new approaches that integrate the effects of  
4 climate change across entire systems.

5

6 Bayden D. Russell<sup>1\*</sup>, Christopher G. D. Harley<sup>2</sup>, Thomas Wernberg<sup>3,4</sup>, Nova Mieszkowska<sup>5</sup>,  
7 Stephen Widdicombe<sup>6</sup>, Jason M. Hall-Spencer<sup>7</sup> and Sean D. Connell<sup>1</sup>.

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9 <sup>1</sup> Southern Seas Ecology Laboratories, School of Earth & Environmental Sciences, The  
10 University of Adelaide, Adelaide, SA, 5005, Australia

11 <sup>2</sup> Biodiversity Research Centre, University of British Columbia, 6270 University Blvd.,  
12 Vancouver, BC, V6T 1Z4, Canada

13 <sup>3</sup> UWA Oceans Institute and School of Plant Biology, University of Western Australia,  
14 Crawley 6009 WA, Australia

15 <sup>5</sup> Marine Biological Association of the UK, The Laboratory, Citadel Hill, Plymouth PL1 2PB,  
16 UK

17 <sup>6</sup> Plymouth Marine Laboratory, Prospect Place, West Hoe, Plymouth PL13DH United  
18 Kingdom

19 <sup>7</sup> Marine Institute, Marine Biology and Ecology Research Centre, University of Plymouth,  
20 Plymouth PL4 8AA, UK

21

22 \* Author for correspondence (bayden.russell@adelaide.edu.au)

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24

25 **SUMMARY**

26 Most studies that forecast the ecological consequences of climate change target a single  
27 species and a single life stage. Depending on climatic impacts on other life stages and on  
28 interacting species, however, the results from simple experiments may not translate into  
29 accurate predictions of future ecological change. Research needs to move beyond simple  
30 experimental studies and environmental envelope projections for single species towards  
31 identifying where ecosystem change is likely to occur and the drivers for this change. For this  
32 to happen, we advocate research directions that (1) identify the critical species within the  
33 target ecosystem, and the life stage(s) most susceptible to changing conditions, and (2) the  
34 key interactions between these species and components of their broader ecosystem. A  
35 combined approach utilising macroecology, experimentally derived data and modelling that  
36 incorporates energy budgets in life-cycle models may identify critical abiotic conditions that  
37 disproportionately alter important ecological processes under forecasted climates.

38

39 **Keywords:** climate change; ocean acidification; global warming; species interactions;  
40 ecosystem shift; productivity and consumption.

41

42

43 **1. INTRODUCTION**

44 The role of global environmental change in altering marine ecosystems has received  
45 increasing attention over the past decade. Global sea surface temperatures have been  
46 warming at  $\sim 0.13^{\circ}\text{C}$  per decade since the current period of climate warming began in the  
47 mid-1980s [1]. Further, marine waters have absorbed  $\sim 30\%$  of  $\text{CO}_2$  emissions and many  
48 marine species are already being forced to cope with increasing ocean acidification in  
49 combination with rising temperatures and other anthropogenic stressors (e.g. eutrophication  
50 and over fishing) [1, 2]. While there is now a substantive body of literature demonstrating  
51 some of the potential negative and positive effects of these combined stressors, the vast  
52 majority of studies currently focus on a single species and life-stage and very few examine  
53 effects on species which play dominant structuring roles in ecosystems (e.g. herbivores, [3];  
54 habitat forming species [4, 5]). Knowledge of the physiological responses of individual  
55 species to environmental change and their limits to performance is an informative first step in  
56 understanding the possible effects of climate change [6]. Extrapolating these physiological  
57 effects on single life history stages of individual species to generalise about changes in  
58 populations or ecosystems is, however, fraught with potentially large forecasting errors  
59 because it fails to take into account two important aspects: (1) the effect of altered  
60 environmental conditions across entire life cycles of the organism; and (2) the interactions of  
61 these species with other components of their ecosystem (e.g. trophic interactions). Yet  
62 experimental manipulations of complete life histories and whole-ecosystems are often  
63 impractical, so an approach which combines experiments and modelling may be necessary.

64

65 To reconcile these issues, a workshop was convened at the University of Plymouth, UK, 28<sup>th</sup>  
66 June – 1<sup>st</sup> July, 2011, to identify gaps in the current research into the role of climate change in  
67 causing ecosystem shifts, how these shifts may be countered by adaptation of plants and

68 animals, and to set future directions for linking seemingly disparate fields of research (e.g.  
69 physiology and macroecology). The workshop included a selection of international specialists  
70 spanning plant and animal physiology, experimental and broad-scale ecology, and ecosystem  
71 modelling.

72

## 73 **2. INTEGRATING INFORMATION ACROSS LIFE STAGES**

### 74 **(a) *Empirical experiments***

75 Understandably, most experimental studies to date have focused on the most easily  
76 manipulated life stage of species, usually mature adults, to quantify physiological changes  
77 and early life stages (e.g. larvae and spores) for growth and development. However, adult  
78 stages often respond differently to earlier life stages and either or both, may be responsible  
79 for regulating population growth and equilibrium population size. For example, it may be of  
80 limited predictive value to detect minor effects of increasing temperature on the adult stage of  
81 a species if it has higher thermal tolerances and/or lower body temperatures than the juvenile  
82 stage [e.g. 7]. Conversely, altered mortality of the early life-stages may be trivial if  
83 recruitment rates are more than sufficient to saturate adult habitat [e.g. 8].

84

85 In addition to this current narrow focus, the perceived necessity of having significant  
86 biological differences among treatments in order to publish has meant that experimental  
87 conditions are often manipulated to unrealistic levels (e.g. CO<sub>2</sub> of >1500 ppm, acute  
88 temperature gradients >20°C) to detect an effect on the more robust adult life stages. While  
89 such extremes are informative about the tolerance limits of the species in question, their use  
90 neglects to identify smaller biological effects that may have multigenerational effects in  
91 populations. Further, these extreme manipulations may not reflect real changes to conditions  
92 over the next century. For example, mature marine molluscs may survive temperature

93 increases within what is predicted in the next 100 years [9, 10], yet if increased temperatures  
94 within this range cause altered reproductive capacity which is not identified in short-term  
95 experiments, then potentially important population and ecosystem effects may not be  
96 predicted. One way to potentially overcome this issue would be to identify the energy budget  
97 of animals and how they allocate resources to different biological processes. This should then  
98 identify if individuals are changing their allocation of energy to ensure maximum survival in  
99 altered environmental conditions at the expense of, or benefit to, other processes important to  
100 population dynamics, such as gonad development [11].

101

102 **(b) *Multi-life stage models***

103 Identifying the stage in the life cycle which is most susceptible to changing environmental  
104 conditions can be challenging, yet necessary to discover where population effects may occur  
105 and any appropriate management or conservation actions to counter them. Detection of an  
106 effect of predicted future conditions (e.g. increased CO<sub>2</sub> and temperatures) with empirical  
107 experiments does not necessarily demonstrate that a particular life stage is the most  
108 susceptible to these conditions *or* that impacts on this life stage will alter population size  
109 unless experiments are conducted across all of the life stages *and* these life stages are  
110 integrated into a complete life cycle. Demographic population models incorporating all life-  
111 cycle stages, which force different scenarios of environmental conditions, can be useful tools  
112 to identify which life-stages are most susceptible, and how this susceptibility may respond to  
113 different combinations of stressors. For example, time-series data for co-occurring species of  
114 warmwater, coldwater and non-native barnacles in the UK have been used to build population  
115 models which show alternate responses of the species to changing conditions; the coldwater  
116 *Semibalanus balanoides* is directly affected by temperature, with pre-recruitment larvae  
117 being the most susceptible stage, whereas the warmwater *Chthamalus montagui* and *C.*

118 *stellatus* are predominantly controlled by competition for settlement space. Importantly, the  
119 invasive *Austrominius (Elminius) modestus* is least likely to be affected by temperature or  
120 acidification, due to its wide thermal and pH tolerance ranges [12, 13], suggesting that a  
121 community shift is likely under future conditions. We suggest that this comparative approach  
122 between interacting species is one of the next key steps in identifying potential ecosystem  
123 shifts driven by changing environmental conditions.

124

### 125 **3. CASE STUDY**

126 Variation in abundance of ecosystem dominants (e.g. kelp forests, coral reefs) reflects a  
127 balance between rates of primary production and its consumption, and ecosystem shifts may  
128 occur when environmental conditions cause large changes in consumption [e.g. 14, 15, 16] or  
129 production [e.g. 17]. We chose to use temperate rocky reefs naturally dominated by kelp  
130 forests as a case study. For simplicity, we assumed that no new species were introduced to the  
131 system because of changing conditions (c.f. range expansion of herbivores due to warming  
132 and the associated ecosystem shift; [18]). In this system, predictions of phase-shifts from kelp  
133 forests to small filamentous turf-forming algae centre on increased productivity of turfs with  
134 increased CO<sub>2</sub> and temperature [17, 19]. However, metabolic theory predicts that herbivores  
135 should be able to consume this additional primary productivity and biomass [15], thus  
136 enabling the system to resist the phase-shift. On the level of an individual adult herbivore this  
137 may be true [20]. When a stressor is integrated across all stages of the life-cycle, however, a  
138 population-level response may become apparent. For example, adult herbivores, in this case  
139 predominantly molluscs, but also including urchins, may be able to function at their normal  
140 levels under temperatures predicted in the next century. Indeed, they may increase their  
141 consumption to compensate for their responses to increasing stress [21]. In short-term  
142 experiments at this single level of ecosystem organisation, it would appear that herbivores

143 increase ecosystem resistance to elevated temperature and CO<sub>2</sub> by consuming the extra algal  
144 biomass resulting from greater rates of primary productivity. Yet natural long-term  
145 experiments at CO<sub>2</sub> vents show that this is unlikely, as herbivore populations tend to decline  
146 under predicted conditions [21]. Therefore, while initial experiments suggested increased  
147 ecosystem resilience as a result of increasing herbivory, population responses to stressors can  
148 be diminished [21], leading to a reduction in ecosystem resilience and potential phase-shifts  
149 as kelp competitors increase in abundance [17, 22].

150

#### 151 **4. OUTCOMES AND CONCLUSIONS**

152 There is clearly a need for research into the potential effects of climate change to move  
153 beyond studies of single species and towards identifying where ecosystem change is likely to  
154 occur and the drivers for this change. The derivation of conceptual models that can be tested  
155 across multiple coastal systems globally will also help to address the current problem faced  
156 by studies of regime shifts; namely that although detection of past shifts is improving with  
157 the benefit of time-series spanning multiple trophic levels, it is still not possible to predict  
158 when and where future events may occur [23]. For this to happen, we advocate two directions  
159 of research: (1) identifying the critical species within the ecosystem in question, and the life  
160 stage(s) which is most susceptible to changing conditions and (2) the interactions of these  
161 species with other components of their ecosystem (e.g. increased or decreased consumption,  
162 whether individual or population based). A combined approach using macroecology,  
163 manipulative experiments and modelling, incorporating energy budgets in life-cycle models,  
164 may identify points where critical biological processes are strongly altered at predicted future  
165 conditions. Importantly, bringing this group of researchers together from seemingly disparate  
166 fields revealed consensus on the need for the field to progress beyond single species studies.  
167 We advocate that with a combined approach it may be possible to predict likely ecosystem



168 changes before reaching what is currently thought of as critical thresholds that are notoriously  
169 difficult to predict.

170

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175

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